

RIPARIAN BASELINE DATA COLLECTION PRIOR TO OPENING THE PECOS RIVER TO FISHING ACCESS IN PECOS NATIONAL HISTORICAL PARK



Final Report

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Summary

This report presents the results of studies conducted along the Pecos River in Pecos National Historical Park (PECO) undertaken to provide baseline information regarding riparian condition prior to the opening of the river to the general public for fishing. Sixty two transects were evaluated as part of a systematic survey design. Survey points, spaced approximately 80 meters from one another along the channel center line, were evaluated for accessibility and habitat quality, and photo points were recorded to document vegetation structure. Vegetation was evaluated along transects placed perpendicular to stream flow at a subset of sample points. Additional metrics from existing rapid assessment methods were evaluated for use in PECO. Results provide a broad baseline from which to evaluate trends in riparian condition as fishing access increases. The report also presents general recommendations for future monitoring.

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Introduction

The National Park Service (NPS) manages an approximately 3 mile long section of the Pecos River in Pecos National Historical Park. The Pecos River and its associated riparian communities are an important resource regionally and a primary concern for managers. Riparian areas provide a range of essential ecosystem services including erosion control, sediment and nutrient removal, and material and energy inputs to aquatic ecosystems (National Research Council 2002). Particularly in arid and semi-arid regions, they provide critical habitat for a diverse range of plants and wildlife, many found nowhere else (Sabo et al. 2005; Stohlgren et al. 2005).

Alluvial Rivers like the Pecos are naturally in a state of dynamic equilibrium. Their channels migrate within their floodplains, erode banks at the outside of meander bends, and deposit material as point bars on the inside of meander bends (Bennett and Simon 2004; Gurnell et al. 1998). In stable streams, erosion is balanced by deposition of an approximately equivalent amount of sediment on point bars and other areas where sediment loads exceed transport capacity (Leopold et al. 1964; Madej 2001; Wohl 2000). Stream bank erosion typically occurs by fluvial entrainment or through mass movement of material (Harmel et al. 1999) and is primarily influenced by channel hydraulic conditions, discharge, and stream bank composition. A variety of human impacts can influence erosion rates including indirect effects (e.g. changes to contributing watersheds such as dams, diversions, or changes in the amount of impervious surfaces) and direct on-site impacts (e.g. bank destabilization due to human or livestock use)(Patten 1998).

Riparian vegetation provides important protection for stream banks by decreasing the velocity and erosive energy of flood waters. Roots of riparian vegetation bind soil particles together, thus contributing to stream bank stability by mechanically reinforcing soils (Abernethy and Rutherford 2001; Gray and Leiser 1982; Wynn et al. 2004). Disturbance from trampling, whether from humans, wildlife, or livestock, can negatively affect bank stability, directly through erosion and indirectly through reductions in plant cover and vigor (Cole and Spilldie 1998; Flenniken et al. 2001; Kozlowski 1999; Kutiel and Zhevelev 2001). The severity of impacts will vary depending on the duration and intensity of disturbance.

In 1990, when Pecos National Historical Park (PECO) expanded and assumed management of a portion of the Pecos River, the NPS pledged to provide future recreational fishing access. PECO is implementing plans for opening access to the Pecos River for fishing, and the park anticipates that the park's lottery based fishing permit system will be fully utilized. To prepare for the opening of access, PECO, with the support from the NPS Water Resources Division, sought to develop an assessment of current conditions in riparian areas along the Pecos River and to

develop a baseline for future monitoring. This report presents an overview of the basic sampling design, data collection procedures, and results from fieldwork conducted in 2007. In addition, we present some considerations and suggestions for future monitoring of fishing impacts.

Objectives

Our specific objectives in this study included:

Evaluate existing datasets on riparian ecosystems along the Pecos River

Using a combination of GIS analyses and field data collection, develop a framework for sampling riparian resources along the Pecos River

Collect baseline information on the condition of riparian areas along the Pecos River

Develop recommendations for future monitoring of potential impacts to the Pecos River riparian corridor in response to changes in fishing management

Methods

The focus of this report is on riparian characterization work done along the length of the Pecos River flowing through PECO (Figure 1). We used a variety of approaches including the collection of vegetation cover and composition data, qualitative/semi-quantitative rapid assessments, and photo point sampling. GIS analyses were used to support and augment measurements in the field.

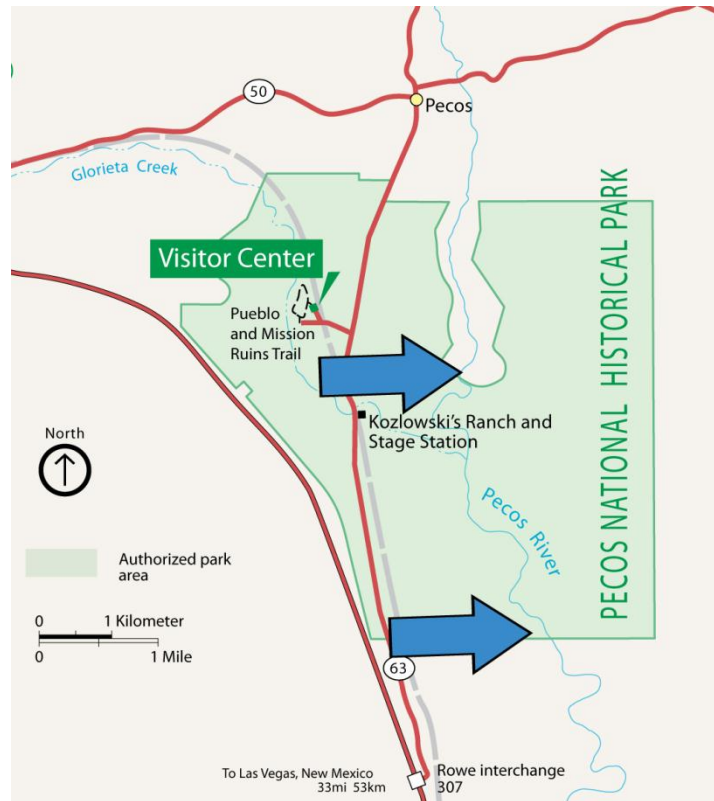


Figure 1. Study area along the Pecos River in Pecos National Historical Park, NM. Work was conducted along an approximately 3 mile stretch of river (bounded between the blue arrows in figure).

As part of baseline data evaluation, we reviewed past work on riparian resources in the park. Habitat characteristics along portions of the Pecos River upstream of the park have been relatively well documented as part of earlier studies and portions of the Pecos River has been sampled as part of broader riparian vegetation classification work conducted within New Mexico (Durkin et al. 1996; Milford et al. 2001; Muldavin et al. 2000; Muldavin et al. 1993; Muldavin 1991). Prior to collecting field data, we reviewed available aerial photographs, National Wetlands Inventory maps, and data products produced by the New Mexico Natural History Program (NMNHP). The riparian and wetland map produced by the NMNHP (Muldavin 1991) and a variety of other existing GIS layers were evaluated for use in stratifying the study area into study reaches. These were useful for evaluating general composition of riparian communities. However, field investigations suggested that there was no reason to assume *a priori* that vegetation type is a good predictor of riparian area response to the type of impacts likely from fishing, as qualities likely to influence the behavior of individual fishermen such as the presence of quality fish habitat features such as pools were not correlated with riparian community types. Accordingly, we decided that a systematic sampling design would provide

better spatial dispersion of sampling points than a stratified random sample drawn from vegetation associations or geomorphic landforms.

Working in ArcGIS, we used a random number as an offset from the intersection of the northernmost property line and the Pecos River, then established 62 points along the channel centerline. Points were equally spaced and located approximately 80 m apart (Figure 2). Sample points were uploaded to a GPS unit as navigation waypoints. All 62 points were visited and photographed and vegetation transects were established in a subset of points.

We expect that the impacts from fishing will be concentrated to a limited number of sites. Therefore, we evaluated stream reaches centered on each sampling point using a simple qualitative scale to assess relative accessibility and fishing habitat quality, two factors we thought would be useful predictors of visitation levels. Separate scores for bank access and fishing habitat quality were assigned to a 100 m long reach centered on each sampling point ($n=62$). Ratings for access and habitat quality were scored as follows: 1=poor, 2=fair, 3=good, 4=excellent. Stream reaches were given low scores for accessibility when either the nature of the terrain or vegetation prevented easy access to streambanks. Fishing habitat quality scores were determined by evaluating the presence of specific habitat features such as pools or overhanging banks. A synthetic metric was created by summing bank accessibility and fishing habitat quality scores, and is intended to provide a gross indicator of sites likely to experience greater traffic and impacts.

At a subset of stream reaches selected for vegetation sampling (Figure 2), a transect was established perpendicular to the direction of river flow, with endpoints typically just outside the boundary between riparian and upland vegetation. Along the stream reaches with very wide riparian areas, sampling was restricted to a zone approximately 12 m wide; the rationale was to focus sampling effort in areas near stream banks that are most likely to be impacted by foot traffic from fishing, and not provide an assessment of all riparian areas. We marked endpoints using rebar driven into the soil and logged the location of each endpoint using a GPS.

Along transects, we made ocular estimates of vascular plant cover by species in a series of contiguous 1x1 m plots centered on the transect line. In addition, we made ocular measurements of the % bare ground, % rock, % wood (branches measuring >3 cm along their intermediate axis), % litter, and the proportion of plot area supporting cryptobiotic crusts or nonvascular plant cover. Plots ran from the transect end to the edge of water. To provide a measure of regeneration, we tallied all seedlings and saplings of woody species.

Two rectangular macroplots, one on each side of the stream, were established along the main transect line. Macroplots had fixed area (100 m²), but varied in plot dimensions based on the length of the baseline; narrower riparian zones (i.e. shorter baselines), had longer plots. These were surveyed for the presence of additional species, the presence of exotic species, and

qualitative evidence of human disturbance. In addition, we measured the diameter at breast height (DBH) of trees >2 m tall and recorded the total number of saplings found in plots. Any evidence of trampling, erosion, or significant disturbance in plots was recorded.

As part of initial pilot work, we evaluated metrics included in existing rapid assessment approaches for assessing riparian condition. For a subset of sampling points, we applied methods presented in Pfankuch (1975), Kershner et al. (2004), Natural Resources Conservation Service's Stream Visual Assessment (SVA), and the Bureau of Land Management's Proper Functioning Condition (Newton et al. 1998; Prichard et al. 1999); however, after initial application to pilot reaches, we concluded that these methods were unlikely to respond to the particular impacts associated with fishing access and were dropped from subsequent sampling.

At all sampling locations, we collected a series of photographs upstream and downstream of the reach following general recommendations presented by Hall (Hall 2001, 2002). Photographs were downloaded to a PC and synced with GPS track logs using GPS photolinker software. This program populates the EXIF header of photographs with geographic information gathered from GPS track logs using time stamped data from both camera and GPS. Vegetation data were entered and checked for accuracy using Microsoft Access 2007. Plant taxonomy follows the USDA PLANTS database (NRCS 2009), although plants were generally keyed to species using the Flora of New Mexico (Martin and Hutchins 1980).

Basic statistical analyses of vegetation and other data were conducted in Minitab. Data used in analyses are provided in Access and Excel format on a data CD accompanying this report. All relevant spatial data, including locations of transects and photo points are archived in ESRI shapefile format along with appropriate metadata, and are included on the CD.

We calculated stream and valley gradients using a 10 m DEM data in ArcGIS 9.2. Sinuosity, calculated as the cumulative length of all line segments in the stream polyline layer, divided by the distance between the start and finish locations, was calculated in a GIS for the length of the Pecos River flowing through Pecos National Historical Park. Using data from the access and fishing quality assessment, we subjectively located 7 cross-sections with high access and habitat quality scores and measured the cross-sectional profile of the stream using a stadia rod and laser level.

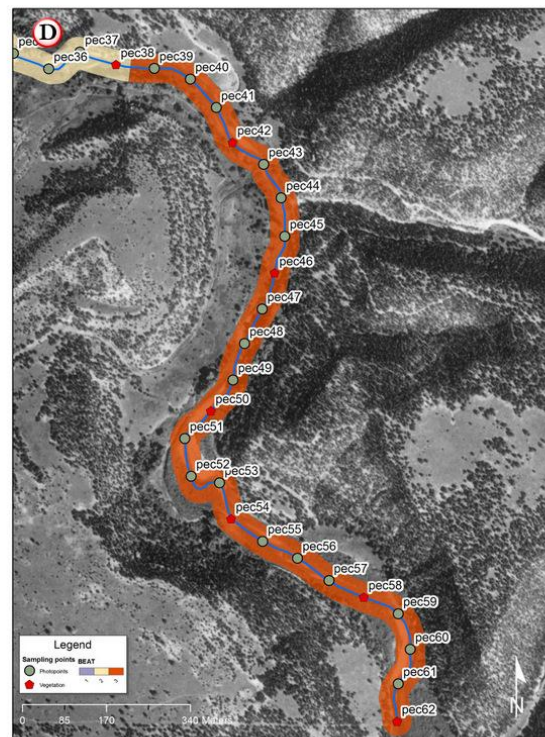
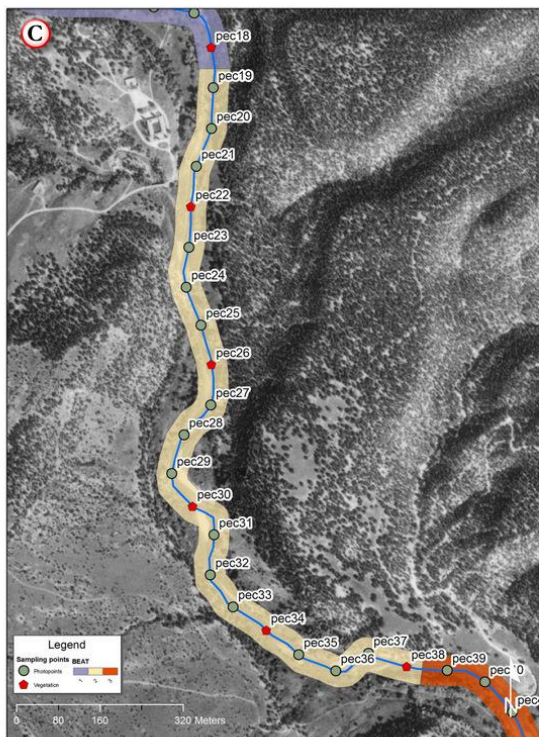
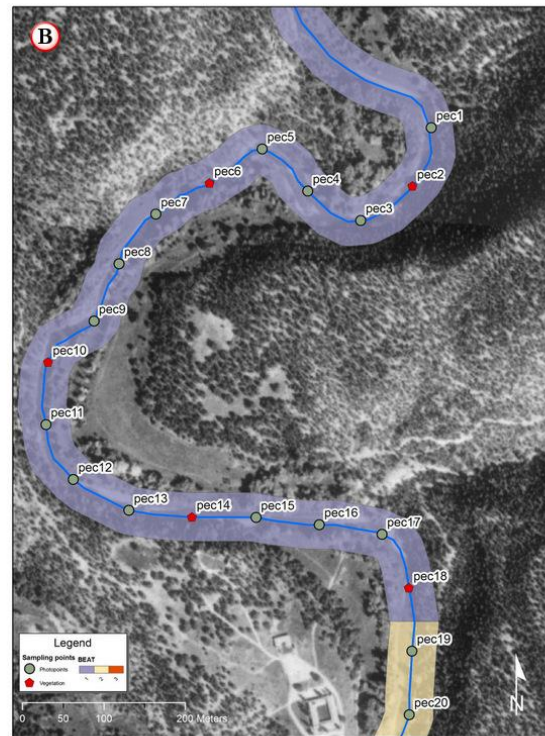
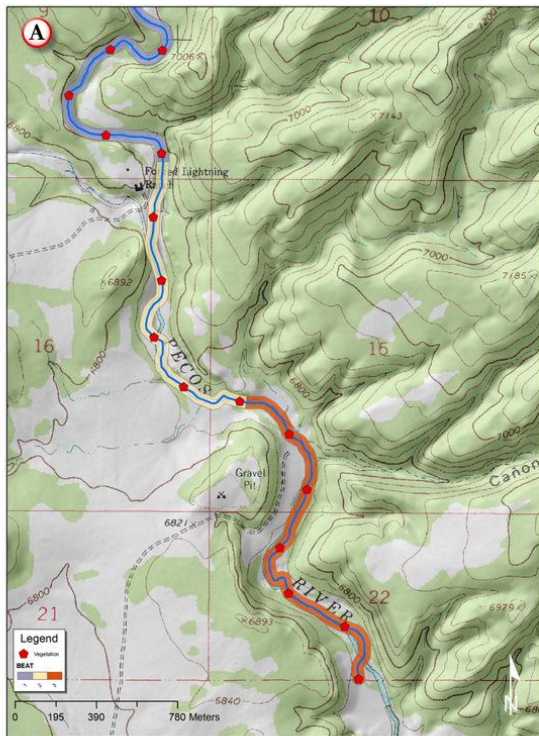


Figure 2. Overview of sampling points along Pecos River (A), and zoomed into the three "beats" (B, C, D) used for fishing permit allocation.

Results

In PECO, the Pecos River is generally only moderately confined and the gradient is lower than steeper canyon reaches upstream of the Park. We calculated a sinuosity of 1.48, a stream gradient of 0.004. The river supports the development of limited mid and side channel bars within an active floodplain. Reaches with long riffles are common, and while less common, pools are distributed throughout the length of the river.

A total of 15 sample points were selected for vegetation analysis, for a total of 30 individual transects (left and right bank). A total of 154 species were recorded in plots (Appendix 2), falling into several different vegetation communities were commonly observed (Table 1). Sampled communities spanned a gradient of water table depth and flood frequency and included obligate hydrophytic species and upland species.

Small patches of hydrophytic vegetation occurred as linear strips adjacent to the channel and within the one to two-year floodplain. Sites supporting these species had fully saturated soils. Other species commonly observed included *Equisetum arvense* (field horsetail) and *Eleocharis parishii* (Parish's spikerush). Small communities dominated by *Typha* spp. (cattails), *Carex* spp., and *Eleocharis* were observed in the wettest sites, but were not typically captured in plots.

Table 1. Vegetation associations described by Muldavin (1993) encountered in transects.

| Vegetation association |
|---|
| Emergent Herbaceous Wetland |
| Coyote Willow Shrub Wetland |
| Rabbitbrush Shrubland |
| Old Field/Disturbed Ground (Upland Toeslope) |
| Narrowleaf Cottonwood Forested Wetland |
| Juniper/Bluegrass Meadow |
| Plains/Narrowleaf Cottonwood forested Wetland |

Communities dominated by *Salix exigua* (Coyote willow) were commonly encountered along transects. These phreatophytic shrubs often co-occur with riparian trees such as *Populus angustifolia* and *Populus x acuminata*. Common understory dominants included *Poa pratensis* (Kentucky bluegrass), *Juncus arcticus*, and a variety of herbaceous dicots. On the drier edges of transects, upland species such as *Juniperus monosperma* (one-seeded juniper) are common.

Riparian forests dominated by *Populus angustifolia* and *P. x acuminata* were common along sampled reaches. *Populus x acuminata*, the hybrid between narrowleaved and plains cottonwood, was the abundant tree species in most stands. *Populus deltoides* was much rarer, but was observed in a handful of locations. Tree cover was often high 90-100%, heavily shading the understory. *Poa pratensis* was

particularly abundant in the understory; other common species included *Erigeron flagellaris* (fleabane) *Melilotus officinalis* (yellow sweetclover).

In the higher portions of floodplains, communities dominated by upland trees and shrubs were encountered. Common understory dominance included such upland species as *Bouteloua gracilis* (blue gramma), *Gutierrezia sarothrae* (snakeweed), *Bouteloua curtipendula* (side-oats gramma), and *Sporobolus spp.* (dropseed). Along the outer margins of the riparian zone, upland communities dominated by *Juniperus monosperma* are common.

Of the 154 species recorded along transects and in plots, 29 species (16.8% of the total list) are listed as non-native in the PLANTS database. Many of these were pasture grasses purposely introduced in the past to provide for forage for livestock. Other common exotics included species like *Phalaris arundanceae* (reed canary grass), *Melilotus officinalis* (yellow sweet clover), and *Taraxacum officinale* (common dandelion).

The portion of plot areas with bare ground varied widely, from <1% to over 50% (Figure 3). Given that likely impacts from fishing will include localized areas of soil disturbance, measures of bare ground may be responsive to disturbance and should be included in future monitoring efforts. Other ground surface characteristics such as % rock or % litter were also variable among plots and transects, although these are less useful for monitoring purposes. Only a small portion of plots in the driest sites had appreciable cover by cryptobiotic crusts or nonvascular plants.

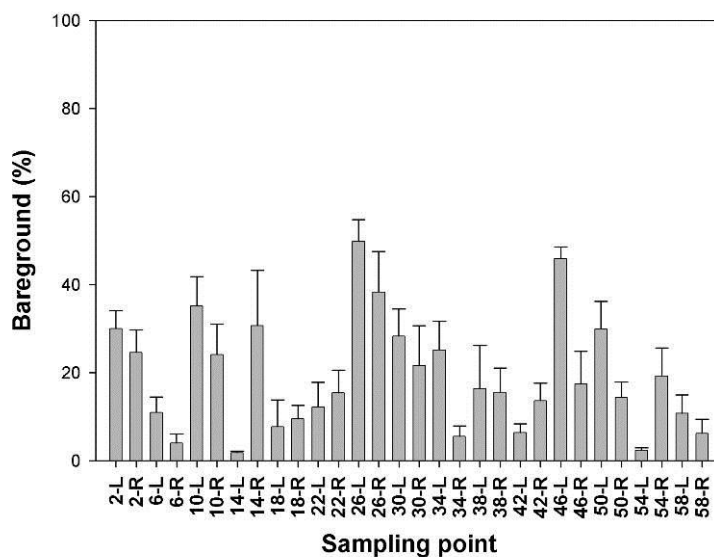


Figure 3. Mean percent bareground (+/- 1 SEM) for vegetation transects.

There was evidence of extensive woody plant regeneration across transects (Figure 4). *Populus angustifolia* seedlings and saplings, for example, were encountered in 60% of transects and associated

macroplots (Figure 4). In general, the frequency of saplings in plots was greater than seedling, although seedlings were commonly observed as well.

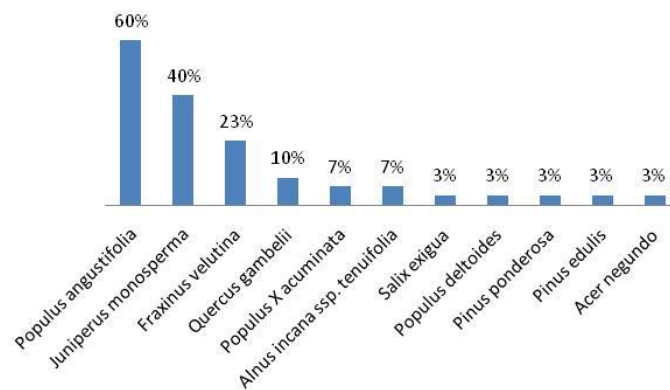


Figure 4. Proportion of total transects sampled (n=30) where tree regeneration was recorded.

Qualitative scores for relative bank access and fishing habit quality were variable. Habitat quality did not significantly differ among the three sections delineated by Park management for administration of fishing permits (i.e. “beats”)(ANOVA, $F=0.53$, $p=0.39$). Access scores were significantly different among reaches (ANOVA, $F=5.35$, $p=0.006$), with Beat #1 having statistically higher access scores than either Beats #2 or #3. Results were similar for the combined scores (ANOVA, $F= 4.87$, $p=0.009$), with Beat #1 having a small, but statistically significant greater score than Beats #2 and #3 (Figure 5).

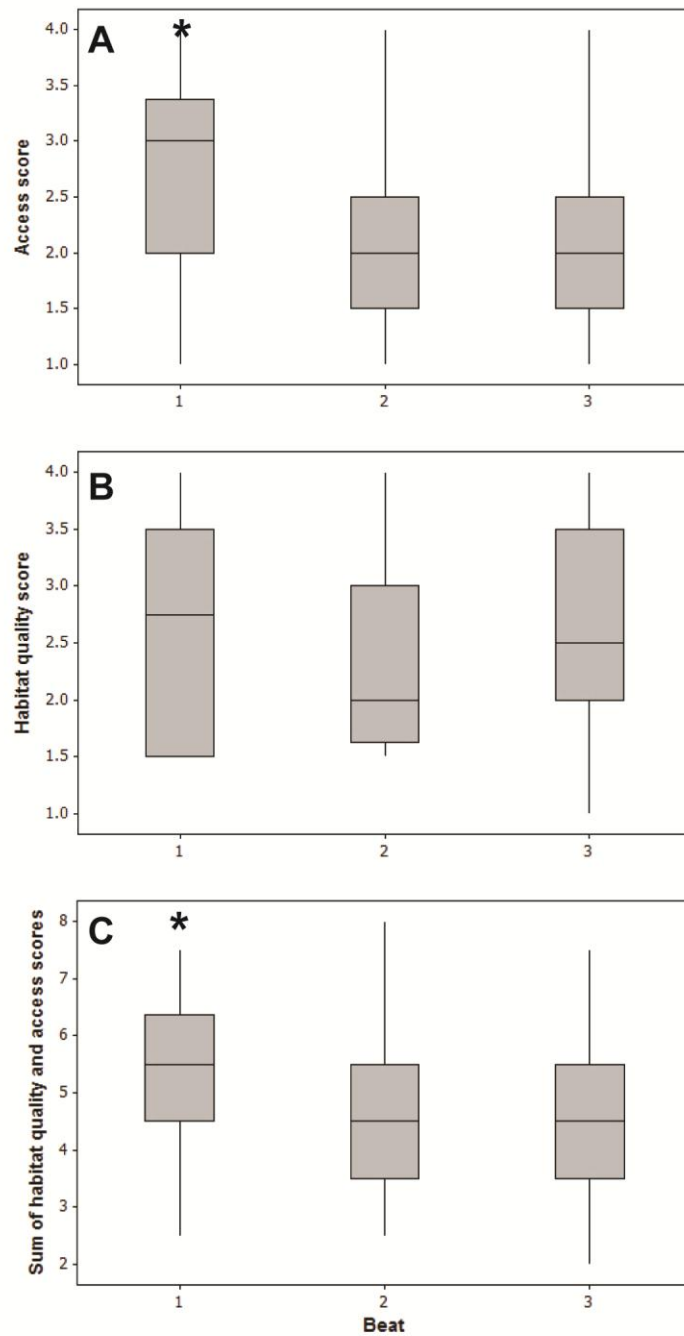


Figure 5. Boxplots of access (A) and fishing habitat quality scores (B), and their sum (C), for all transects (n=62). Rating for starred beats are significantly different ($\alpha=0.05$, ANOVA, Tukey adjustment). Ratings for access and habitat quality are as follows: 1=poor, 2=fair, 3=good, 4=excellent.

Over 500 photographs were taken along sampled reaches (Figure 6). The utility of photos for monitoring is variable. Many reaches, for example those dominated by dense stands of *Salix exigua*, were difficult to effectively photograph.



Figure 6. Example of a series of photopoints taken along a sample reach. All photos have been geotagged and included on a CD as a supplement to the report.

Discussion and recommendations

The Pecos River in PECO supports a diverse and highly functioning riparian corridor. Riparian areas here have been subject to relatively few anthropogenic stressors for decades. Compared to most western rivers of comparable size or to sections of the Pecos River downstream of the park, the Pecos River upstream of PECO has experienced fewer systemic changes. The persistence of natural flow and disturbance regimes helps support a variety of different community types.

However, the river is by no means pristine. Historical cattle grazing has affected plant communities. The riparian corridor support large populations of several non-native plant species, including several pasture grasses purposely introduced as forage for livestock. Changes in the intensity of management and processes of Juniper encroachment have contributed to

changing conditions. Localized areas of bank instability and erosion were observed, although it was generally not clear how much of this was the result of anthropogenic factors.

There are several challenges associated with effective monitoring of riparian resources. First and foremost is the high degree of variability typical of riparian ecosystems. Diverse hydrologic and geomorphic environments, coupled with spatially and temporally variable disturbance processes, help maintain a range of riparian community types. Effective monitoring must be able to detect actual trends in key variables, but not falsely attribute systemic change to natural patterns of variability. Additionally, monitoring objectives must be feasible with available resources, since the most carefully developed monitoring protocol is of little value if it cannot be implemented in the field.

Monitoring should be focused on the kind of impacts most relevant to the key ecological functions of interest and metrics selected should be those most likely to respond to specific agents of change. A variety of monitoring protocols have been developed for riparian assessment. While each has value, most have been developed for different purposes and are focused on different stressors. No existing method is tailored to the kinds of specific impacts likely to occur as a result of increased fishing access.

Impacts from increased public access for fishing can be separated into two general categories. Fishermen may indirectly impact areas by serving as vectors for exotic species. These effects are difficult to capture with monitoring, since many exotic species are already present in PECO riparian areas and it is hard to link changes in the distribution and abundance directly to fishing access. A second category includes direct impacts associated with increases in soil disturbance. Impacts from increases in visitor access will most visibly be manifested in the formation of trails and other eroded areas. These impacts can lead to a variety of ecological as well as aesthetic effects (Figure 7). During fieldwork, few observations of trailing were observed, all attributed to game. Incidences of trailing can be documented opportunistically during routine resource management activities. A few simple notes, photographs and GPS coordinates can provide useful information and need not require excessive commitment by Park staff. However, a more systematic approach to documenting these impacts should be undertaken periodically.

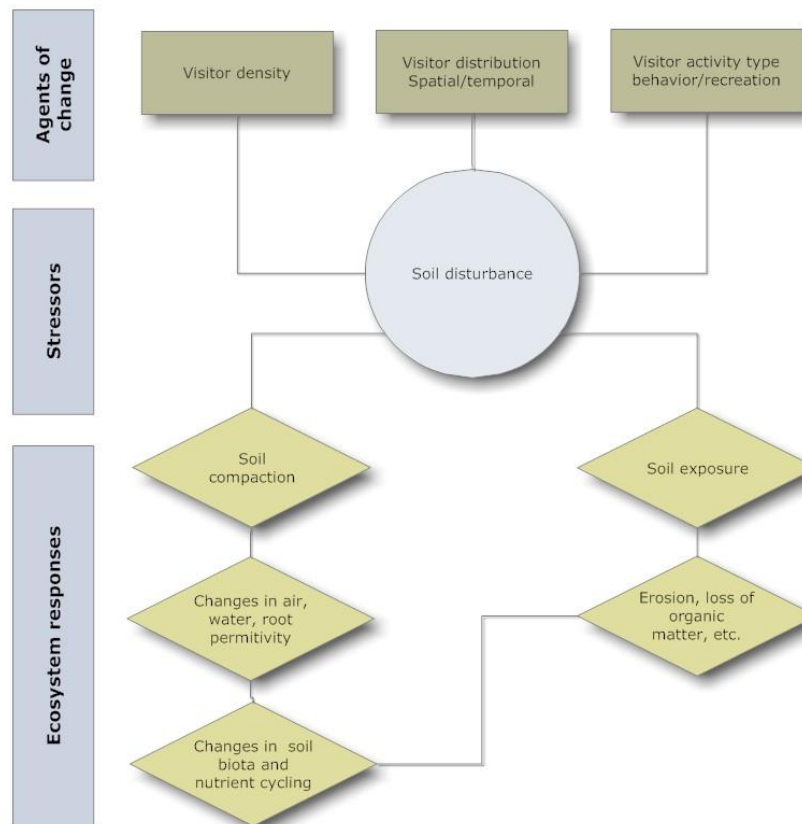


Figure 7. Conceptual model illustrating the kinds of stressors and ecosystem responses that may be expected from an increase in soil disturbance accompanying increased visitor use (modified from Monz and Leung 2006). Note that this model deals with soil disturbance only; other impacts such the spread of exotic species may also occur with increased public access.

In our evaluation of existing rapid assessment protocols commonly used to evaluate wetland and riparian condition and function, we found that all have elements useful to riparian monitoring in PECO, but taken as a whole, no single approach is very well suited to the specific kinds of impacts likely from increased fishing. A key consideration in selecting metrics for monitoring is precision. All methods are subject to some inter-observer error; if this is too large, any meaningful trend information is lost, or trend can be falsely detected where there is none. A general weakness of most qualitative or semi-quantitative methods is that the specific metrics are so vague as to have different meanings to different individuals. To be really useful, all methods need to be precisely qualified with respect to location. Even if different observers make identical calls regarding the value of specific metrics, their outcomes will differ if they use different assessment areas (e.g. a 20 m versus 300 m stream reach).

Overall, Pfankuch's (1975) Stream Reach Inventory and Channel Stability Evaluation method is more appropriate for broader scale assessments and comparative assessments (e.g. different basins and streams). We found streambank stability indicators included in Kershner et al. (2004) were conceptually useful, but as a baseline for monitoring fishing impacts, they were too

general to recommend for monitoring purposes. Attributes such as the presence of slump blocks or cracks in the streambank could indicate impacts from trampling, but also occur from natural erosional processes, so absent a specific frame of reference, which may not overlap with areas where fishing impacts are concentrated, the approach is not likely useful for trend detection.

The Natural Resources Conservation Service's Stream Visual Assessment (SVA), and Bureau of Land Management's Proper Functioning Condition (PFC) assessment were evaluated for possible inclusion with recommended monitoring protocols (Prichard et al. 1998; Ward et al. 2003). Metrics included with both the SVA and PFC methods were generally found to lack sufficient precision to be particularly useful for monitoring specific impacts from increased access for fishing. Neither method is likely sensitive enough to detect the localized impacts most likely to occur. Some attributes included in these methods such as the presence of undercut banks, slump blocks (i.e. partially or completely detached pieces of the streambank), are useful and can be included in general assessment and observation reports made by park staff.

A key objective of this research was to collect baseline data and recommend potential monitoring approaches for evaluating potential impacts of opening fishing access to the Pecos River. There are a significant number of factors that affect the ability of any particular monitoring strategy to detect and measure change. Limitations in resources for sampling must also be frankly recognized. As a consequence, the rather labor intensive methods used for baseline data collection are not appropriate for routine annual monitoring. The availability of monitoring personnel, particularly individuals with botanical training, is limited, so approaches should emphasize rapid assessment techniques such as photo points and semi-quantitative or qualitative assessments. The more extensive and intensive nature of sampling used for baseline data collection can be episodically revisited as resources become available.

Several characteristics make sampling likely impacts from fishing difficult. Impacts will tend to be highly localized and non-random, influenced by a variety of factors. The relative ease of travel, as influenced by the terrain (slopes, cliff-sections) and vegetation (dense willow, juniper), will tend to localize traffic in certain areas. Proximity to access points will also influence the spatial pattern of impacts. Presumably, sites closer to access points will receive greater use, and therefore will be more likely to experience measurable impacts. High water, early in the season will prevent most people from accessing the opposite banks during late spring and early summer. Fisherman behavior is also influenced by skill level, motivation, and equipment (e.g. waders), which may affect patterns of use and impact. For example, banks near large pools will likely see more traffic than densely wooded riffle sections difficult to access from the bank.

Monitoring is most effective when it can balance precision, accuracy, and efficiency. Likely impacts from fishing are dispersed and complex, making effective monitoring difficult given likely constraints on resources for sampling. The relationship between visitor use and impacts is often curvilinear, with much of the impact occurring with initial use (Cole and Monz 2003; Monz and Leung 2006). This, coupled with the fact that the area has been closed to public access for so long, suggests that many problems will become apparent within the first couple years of use. Currently, the riparian corridor is largely free of trails and other visible signs of disturbance; this will inevitably change over time. We recommend that developing social trail networks should be periodically inventoried and mapped using a GIS (Marion et al. 2006). This information will allow Park managers monitor changing conditions and help define acceptable levels of impact based on on-the-ground observations.

Regardless of specific monitoring approaches adopted, management should be adaptive to conditions on the ground. Considering that the riparian area in PECO has been closed to public access for so long, it is inevitable that some minor trailing will occur, but most instances will not likely be actionable by Park managers because no significant resource damage is occurring. However, there may be areas where trailing leads to more significant impacts (e.g. causing significant hill slope erosion), necessitating some management action. Regular monitoring, even if only consisting of general reconnaissance of riparian areas, is important so these problem areas can be identified early. Whenever feasible, resource management staff should carry a GPS unit and camera; potential problem areas observed during unrelated job activities may be documented using geo-tagged photo points and flagged for follow up.

Good information regarding the spatial and temporal patterns of visitor use along the Pecos River can facilitate more efficient and effective use of limited monitoring resources. Many areas will see little use, while areas near access points may see high levels of use. By monitoring use patterns, problem areas can be identified early. The frequency of visits (i.e. visitor numbers) and the intensity of use will be one of the strongest correlates of resource impacts.

Basic information will be collected as part of the administration of the fishing permits. However, because fishing impacts will tend to be localized, absent more specific information on the spatial pattern of visitor use, these data will be too coarse to help locate potential problem areas. Several options exist for augmenting visitor use information, providing more effective ways of concentrating management efforts.

A variety of trail counters on the market are routinely used by recreation managers in National Parks and National Forests to develop recreational use statistics. Counters can be placed systematically along the river to help refine understanding of human traffic or in strategic locations where traffic naturally consolidates. Scouting cameras can be used in a similar fashion, and over the long-term, can be a cost-effective tool to monitor use. Both require little maintenance and can be revisited infrequently, minimizing time requirements by park staff.

GPS loggers can be an effective tool to develop an understanding of the spatial pattern of visitor use. Many low-cost data logging GPS transponders are on the market. These lack all of the navigation and other user features found in more sophisticated units, but can be purchased very inexpensively (<\$60). It is probably not feasible to require visitors to carry these; many visitors might object for reasons such as privacy. However, the park can incentivize users by presenting clear information on how the data will be used and by offering a fee discount or waiver. Data from even a small fraction of total visits will still provide a much more refined picture of use patterns along the river.

Procedures used to sample vegetation were time-consuming and do not represent a viable approach for annual monitoring by parks staff. However, these data can be revisited less frequently, perhaps with the help of external resources. These more rigorous and quantitative approaches to monitoring complement information collected using more rapid approaches such as photopoint monitoring.

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Appendix 1. List and map of sampling points

| Reach | UTM -X | UTM-Y | Beat | Sampling | Reach | UTM -X | UTM-Y | Beat | Sampling |
|-------|-----------|------------|------|------------|-------|-----------|------------|------|------------|
| 1 | 439470.63 | 3933341.60 | 1 | Photopoint | 32 | 439440.94 | 3931754.03 | 2 | Photopoint |
| 2 | 439447.37 | 3933269.41 | 1 | Vegetation | 33 | 439484.81 | 3931692.14 | 2 | Photopoint |
| 3 | 439383.38 | 3933226.86 | 1 | Photopoint | 34 | 439549.24 | 3931647.15 | 2 | Vegetation |
| 4 | 439318.04 | 3933263.23 | 1 | Photopoint | 35 | 439611.46 | 3931599.74 | 2 | Photopoint |
| 5 | 439261.83 | 3933315.37 | 1 | Photopoint | 36 | 439683.00 | 3931568.36 | 2 | Photopoint |
| 6 | 439197.06 | 3933273.28 | 1 | Vegetation | 37 | 439746.52 | 3931602.56 | 2 | Photopoint |
| 7 | 439130.45 | 3933234.63 | 1 | Photopoint | 38 | 439820.02 | 3931577.18 | 2 | Vegetation |
| 8 | 439084.94 | 3933173.45 | 1 | Photopoint | 39 | 439898.17 | 3931569.33 | 3 | Photopoint |
| 9 | 439053.97 | 3933102.40 | 1 | Photopoint | 40 | 439971.07 | 3931547.31 | 3 | Photopoint |
| 10 | 438997.37 | 3933051.82 | 1 | Vegetation | 41 | 440023.93 | 3931489.20 | 3 | Photopoint |
| 11 | 438994.65 | 3932974.55 | 1 | Photopoint | 42 | 440057.66 | 3931418.69 | 3 | Vegetation |
| 12 | 439028.33 | 3932906.66 | 1 | Photopoint | 43 | 440120.47 | 3931373.79 | 3 | Photopoint |
| 13 | 439096.96 | 3932868.45 | 1 | Photopoint | 44 | 440156.01 | 3931305.25 | 3 | Photopoint |
| 14 | 439175.26 | 3932860.44 | 1 | Vegetation | 45 | 440163.13 | 3931227.20 | 3 | Photopoint |
| 15 | 439254.01 | 3932859.65 | 1 | Photopoint | 46 | 440142.71 | 3931152.76 | 3 | Vegetation |
| 16 | 439332.20 | 3932850.81 | 1 | Photopoint | 47 | 440117.29 | 3931078.87 | 3 | Photopoint |
| 17 | 439409.64 | 3932838.93 | 1 | Photopoint | 48 | 440081.45 | 3931008.77 | 3 | Photopoint |
| 18 | 439442.99 | 3932772.84 | 1 | Vegetation | 49 | 440058.04 | 3930934.59 | 3 | Photopoint |
| 19 | 439446.67 | 3932694.69 | 2 | Photopoint | 50 | 440013.09 | 3930871.69 | 3 | Vegetation |
| 20 | 439443.37 | 3932615.98 | 2 | Photopoint | 51 | 439959.81 | 3930815.46 | 3 | Photopoint |
| 21 | 439413.94 | 3932542.99 | 2 | Photopoint | 52 | 439972.82 | 3930738.32 | 3 | Photopoint |
| 22 | 439403.43 | 3932465.33 | 2 | Vegetation | 53 | 440030.65 | 3930724.99 | 3 | Photopoint |
| 23 | 439399.57 | 3932386.71 | 2 | Photopoint | 54 | 440054.58 | 3930651.80 | 3 | Vegetation |
| 24 | 439394.46 | 3932309.62 | 2 | Photopoint | 55 | 440118.64 | 3930606.03 | 3 | Photopoint |
| 25 | 439423.02 | 3932236.28 | 2 | Photopoint | 56 | 440189.06 | 3930571.33 | 3 | Photopoint |
| 26 | 439443.41 | 3932160.28 | 2 | Vegetation | 57 | 440253.29 | 3930525.92 | 3 | Photopoint |
| 27 | 439441.80 | 3932082.26 | 2 | Photopoint | 58 | 440324.28 | 3930492.18 | 3 | Vegetation |
| 28 | 439390.40 | 3932024.31 | 2 | Photopoint | 59 | 440394.12 | 3930458.15 | 3 | Photopoint |
| 29 | 439366.79 | 3931949.80 | 2 | Photopoint | 60 | 440418.54 | 3930385.96 | 3 | Photopoint |
| 30 | 439407.04 | 3931886.33 | 2 | Vegetation | 61 | 440394.10 | 3930315.66 | 3 | Photopoint |
| 31 | 439448.10 | 3931831.21 | 2 | Photopoint | 62 | 440392.35 | 3930239.41 | 3 | Vegetation |

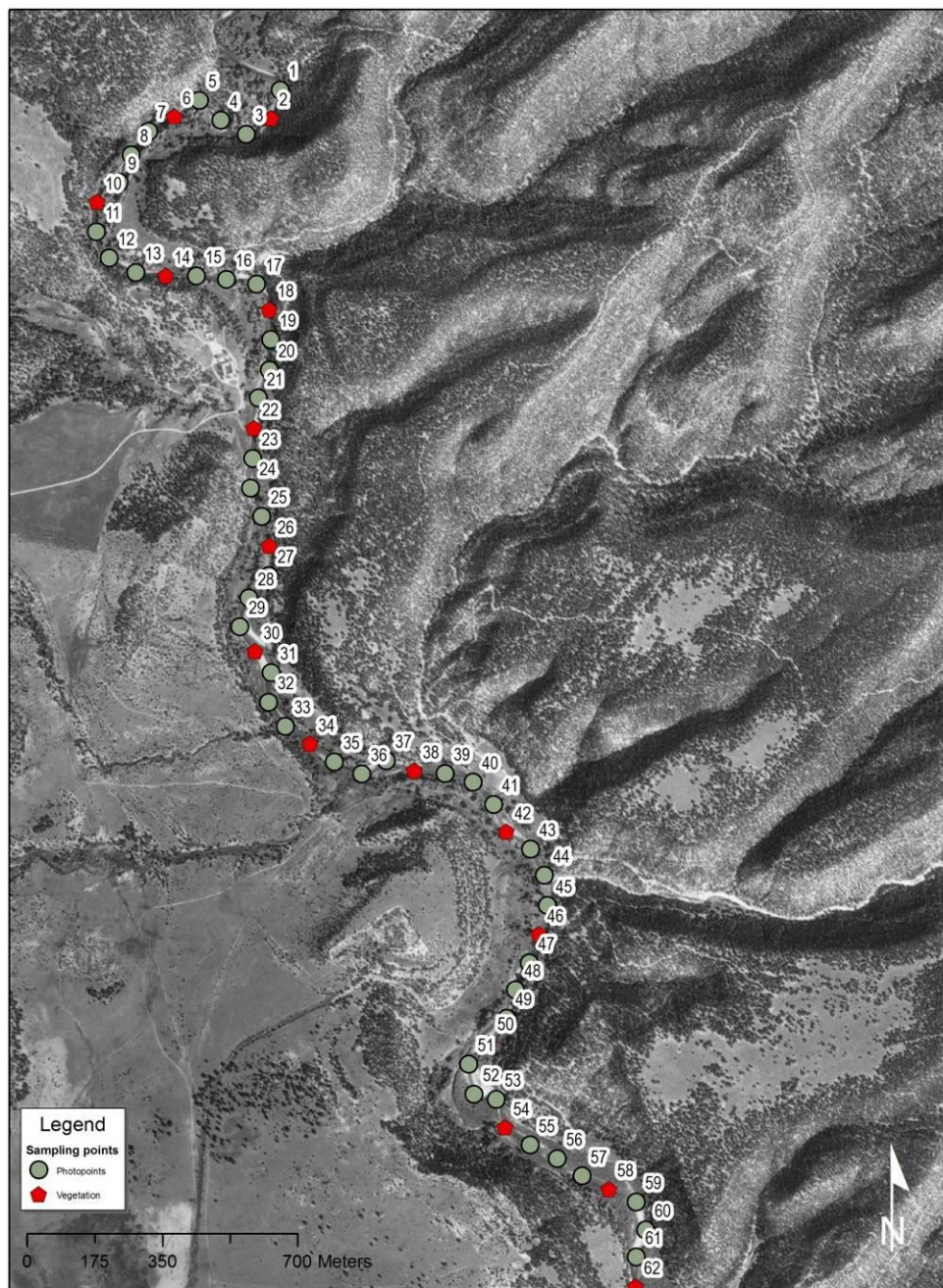


Figure 8. Locations of systematically placed sample reaches along the Pecos River.

Appendix 2. Plant list

Species recorded in vegetation plots. Species in bold are non-natives. Nomenclature follows the USDA PLANTS database.

| species | Common Name | Accepted Symbol | Family |
|--|-------------------------|-----------------|----------------|
| <i>Achillea millefolium</i> | common yarrow | ACMI2 | Asteraceae |
| <i>Actaea rubra</i> | red baneberry | ACRU2 | Ranunculaceae |
| <i>Agrostis exarata</i> | spike bentgrass | AGEX | Poaceae |
| <i>Agrostis gigantea</i> | redtop | AGGI2 | Poaceae |
| <i>Agrostis scabra</i> | rough bentgrass | AGSC5 | Poaceae |
| <i>Allium geoyeri</i> | Geyer's onion | ALGE | Liliaceae |
| <i>Alnus incana</i> ssp. <i>tenuifolia</i> | thinleaf alder | ALINT | Betulaceae |
| <i>Ambrosia acanthicarpa</i> | flatspine bur ragweed | AMAC2 | Asteraceae |
| <i>Ambrosia artemisiifolia</i> | annual ragweed | AMAR2 | Asteraceae |
| <i>Androsace septentrionalis</i> | pygmyflower rockjasmine | ANSE4 | Primulaceae |
| <i>Apocynum cannabinum</i> | Indianhemp | APCA | Apocynaceae |
| <i>Arabis fendleri</i> var. <i>fendleri</i> | Fendler's rockcress | ARFEF | Brassicaceae |
| <i>Aristida purpurea</i> var. <i>longiseta</i> | Fendler threeawn | ARPUL | Poaceae |
| <i>Artemisia carruthii</i> | Carruth's sagewort | ARCA14 | Asteraceae |
| <i>Artemisia frigida</i> | prairie sagewort | ARFR4 | Asteraceae |
| <i>Artemisia ludoviciana</i> | white sagebrush | ARLU | Asteraceae |
| <i>Asclepias speciosa</i> | showy milkweed | ASSP | Asclepiadaceae |
| <i>Astragalus missouriensis</i> | Missouri milkvetch | ASMI10 | Fabaceae |
| <i>Bahia dissecta</i> | ragleaf bahia | BADI | Asteraceae |
| <i>Berberis fendleri</i> | Colorado barberry | BEFE | Berberidaceae |
| <i>Bouteloua curtipendula</i> | sideoats grama | BOCU | Poaceae |
| <i>Bouteloua gracilis</i> | blue grama | BOGR2 | Poaceae |
| <i>Bromus arvensis</i> | field brome | BRAR5 | Poaceae |
| <i>Bromus inermis</i> | smooth brome | BRIN2 | Poaceae |
| <i>Bromus lanatipes</i> | woolly brome | BRLA6 | Poaceae |
| <i>Bromus tectorum</i> | cheatgrass | BRTE | Poaceae |
| <i>Carex emoryi</i> | Emory's sedge | CAEM2 | Cyperaceae |
| <i>Carex occidentalis</i> | western sedge | CAOC2 | Cyperaceae |
| <i>Chenopodium album</i> | lambsquarters | CHAL7 | Chenopodiaceae |

| | | | |
|---|-----------------------------|--------------|------------------------|
| <i>Chenopodium leptophyllum</i> | narrowleaf goosefoot | CHLE4 | Chenopodiaceae |
| <i>Cichorium intybus</i> | chicory | CIIN | Asteraceae |
| <i>Cicuta maculata</i> | spotted water hemlock | CIMA2 | Apiaceae |
| <i>Cirsium arvense</i> | Canada thistle | CIAR4 | Asteraceae |
| <i>Cirsium ochrocentrum</i> | yellowspine thistle | CIOC2 | Asteraceae |
| <i>Clematis ligusticifolia</i> | western white clematis | CLLI2 | Ranunculaceae |
| <i>Conium maculatum</i> | poison hemlock | COMA2 | Apiaceae |
| <i>Conyza canadensis</i> | Canadian horseweed | COCA5 | Asteraceae |
| <i>Crataegus erythropoda</i> | cerro hawthorn | CRER | Rosaceae |
| <i>Dactylis glomerata</i> | orchardgrass | DAGL | Poaceae |
| <i>Dianthus armeria</i> | Deptford pink | DIAR | Caryophyllaceae |
| <i>Dipsacus fullonum</i> | Fuller's teasel | DIFU2 | Dipsaceae |
| <i>Eleocharis palustris</i> | common spikerush | ELPA3 | Cyperaceae |
| <i>Eleocharis parishii</i> | Parish's spikerush | ELPA4 | Cyperaceae |
| <i>Elymus canadensis</i> | Canada wildrye | ELCA4 | Poaceae |
| <i>Elymus trachycaulus</i> | slender wheatgrass | ELTR7 | Poaceae |
| <i>Epilobium ciliatum</i> | fringed willowherb | EPCI | Onagraceae |
| <i>Equisetum arvense</i> | field horsetail | EQAR | Equisetaceae |
| <i>Ericameria nauseosa</i> | rubber rabbitbrush | ERNA10 | Asteraceae |
| <i>Erigeron divergens</i> | spreading fleabane | ERDI4 | Asteraceae |
| <i>Erigeron flagellaris</i> | trailing fleabane | ERFL | Asteraceae |
| <i>Erigeron formosissimus</i> | beautiful fleabane | ERFO3 | Asteraceae |
| <i>Euphorbia brachycera</i> | horned spurge | EUBR | Euphorbiaceae |
| <i>Euphorbia davidii</i> | David's spurge | EUDA5 | Euphorbiaceae |
| <i>Fraxinus velutina</i> | velvet ash | FRVE2 | Oleaceae |
| <i>Gaillardia aristata</i> | common gaillardia | GAAR | Asteraceae |
| <i>Gaillardia pinnatifida</i> | red dome blanketflower | GAPI | Asteraceae |
| <i>Gaura coccinea</i> | scarlet beeblossom | GACO5 | Onagraceae |
| <i>Gaura mollis</i> | velvetweed | GAMO5 | Onagraceae |
| <i>Geranium atropurpureum</i> var. <i>atropurpureum</i> | western purple cranesbill | GEATA | Geraniaceae |
| <i>Geranium caespitosum</i> | pineywoods geranium | GECA3 | Geraniaceae |
| <i>Glyceria grandis</i> | American mannagrass | GLGR | Poaceae |
| <i>Glycyrrhiza lepidota</i> | American licorice | GLLE3 | Fabaceae |
| <i>Gutierrezia sarothrae</i> | broom snakeweed | GUSA2 | Asteraceae |
| <i>Hedeoma drummondii</i> | Drummond's false pennyroyal | HEDR | Lamiaceae |

| | | | |
|---|---------------------------|-------------|---------------------|
| <i>Helianthus annuus</i> | common sunflower | HEAN3 | Asteraceae |
| <i>Helianthus petiolaris</i> | prairie sunflower | HEPE | Asteraceae |
| <i>Hesperostipa comata</i> ssp. <i>comata</i> | needle and thread | HECOC8 | Poaceae |
| <i>Hesperostipa neomexicana</i> | New Mexico feathergrass | HENE5 | Poaceae |
| <i>Heterotheca villosa</i> | hairy false goldenaster | HEVI4 | Asteraceae |
| <i>Hordeum jubatum</i> | foxtail barley | HOJU | Poaceae |
| <i>Ipomopsis aggregata</i> ssp. <i>formosissima</i> | scarlet gilia | IPAGF | Polemoniaceae |
| <i>Ipomopsis longiflora</i> | flaxflowered ipomopsis | IPLO2 | Polemoniaceae |
| <i>Juncus arcticus</i> | arctic rush | JUAR2 | Juncaceae |
| <i>Juncus bufonius</i> | toad rush | JUBU | Juncaceae |
| <i>Juncus tenuis</i> | poverty rush | JUTE | Juncaceae |
| <i>Juncus torreyi</i> | Torrey's rush | JUTO | Juncaceae |
| <i>Juncus tracyi</i> | Tracy's rush | JUTR | Juncaceae |
| <i>Juniperus communis</i> | common juniper | JUCO6 | Cupressaceae |
| <i>Juniperus monosperma</i> | oneseed juniper | JUMO | Cupressaceae |
| <i>Lactuca serriola</i> | prickly lettuce | LASE | Asteraceae |
| <i>Leucanthemum vulgare</i> | oxeye daisy | LEVU | Asteraceae |
| <i>Listera cordata</i> | heartleaf twayblade | LICO6 | Orchidaceae |
| <i>Lygodesmia juncea</i> | rush skeletonplant | LYJU | Asteraceae |
| <i>Machaeranthera canescens</i> | hoary tansyaster | MACA2 | Asteraceae |
| <i>Machaeranthera pinnatifida</i> ssp. <i>pinnatifida</i> | lacy tansyaster | MAPIP | Asteraceae |
| <i>Medicago lupulina</i> | black medick | MELU | Fabaceae |
| <i>Medicago sativa</i> | alfalfa | MESA | Fabaceae |
| <i>Melampodium leucanthum</i> | plains blackfoot | MELE2 | Asteraceae |
| <i>Melilotus officinalis</i> | yellow sweetclover | MEOF | Fabaceae |
| <i>Mentha arvensis</i> | wild mint | MEAR4 | Lamiaceae |
| <i>Monarda fistulosa</i> | wild bergamot | MOFI | Lamiaceae |
| <i>Monarda pectinata</i> | pony beebalm | MOPE | Lamiaceae |
| <i>Muhlenbergia asperifolia</i> | scratchgrass | MUAS | Poaceae |
| <i>Muhlenbergia montana</i> | mountain muhly | MUMO | Poaceae |
| <i>Muhlenbergia torreyi</i> | ring muhly | MUTO2 | Poaceae |
| <i>Nasturtium officinale</i> | watercress | NAOF | Brassicaceae |
| <i>Oenothera albicaulis</i> | whitest evening primrose | OEAL | Onagraceae |
| <i>Oenothera caespitosa</i> | tufted evening primrose | OECA10 | Onagraceae |
| <i>Opuntia phaeacantha</i> | tulip pricklypear | OPPH | Cactaceae |

| | | | |
|--|----------------------------|--------------|------------------------|
| <i>Oxalis alpina</i> | alpine woodsorrel | OXAL2 | Oxalidaceae |
| <i>Parthenocissus vitacea</i> | woodbine | PAVI5 | Vitaceae |
| <i>Pascopyrum smithii</i> | western wheatgrass | PASM | Poaceae |
| <i>Pennellia longifolia</i> | longleaf mock thelypody | PELO3 | Brassicaceae |
| <i>Phalaris arundinacea</i> | reed canarygrass | PHAR3 | Poaceae |
| <i>Phleum pratense</i> | timothy | PHPR3 | Poaceae |
| <i>Phlox nana</i> | Santa Fe phlox | PHNA2 | Polemoniaceae |
| <i>Pinus edulis</i> | twoneedle pinyon | PIED | Pinaceae |
| <i>Pinus ponderosa</i> | ponderosa pine | PIPO | Pinaceae |
| <i>Plantago lanceolata</i> | narrowleaf plantain | PLLA | Plantaginaceae |
| <i>Plantago major</i> | common plantain | PLMA2 | Plantaginaceae |
| <i>Plantago patagonica</i> | woolly plantain | PLPA2 | Plantaginaceae |
| <i>Poa pratensis</i> | Kentucky bluegrass | POPR | Poaceae |
| <i>Polygonum aviculare</i> | prostrate knotweed | POAV | Polygonaceae |
| <i>Populus angustifolia</i> | narrowleaf cottonwood | POAN3 | Salicaceae |
| <i>Populus deltoides</i> | eastern cottonwood | PODE3 | Salicaceae |
| <i>Populus X acuminata</i> | lanceleaf cottonwood | POAC5 | Salicaceae |
| <i>Potentilla pensylvanica</i> | Pennsylvania cinquefoil | POPE8 | Rosaceae |
| <i>Prunella vulgaris</i> | common selfheal | PRVU | Lamiaceae |
| <i>Quercus gambelii</i> | Gambel oak | QUGA | Fagaceae |
| <i>Ranunculus cymbalaria</i> | alkali buttercup | RACY | Ranunculaceae |
| <i>Rhus trilobata</i> | skunkbush sumac | RHTR | Anacardiaceae |
| <i>Rosa woodsii</i> | Woods' rose | ROWO | Rosaceae |
| <i>Rumex acetosella</i> | common sheep sorrel | RUAC3 | Polygonaceae |
| <i>Rumex crispus</i> | curly dock | RUCR | Polygonaceae |
| <i>Salix bebbiana</i> | Bebb willow | SABE2 | Salicaceae |
| <i>Salix exigua</i> | narrowleaf willow | SAEX | Salicaceae |
| <i>Salix gooddingii</i> | Goodding's willow | SAGO | Salicaceae |
| <i>Salix irrorata</i> | dewystem willow | SAIR | Salicaceae |
| <i>Salix lucida</i> | shining willow | SALU | Salicaceae |
| <i>Saponaria officinalis</i> | bouncingbet | SAOF4 | Caryophyllaceae |
| <i>Schizachyrium scoparium</i> | little bluestem | SCSC | Poaceae |
| <i>Schoenoplectus acutus</i> | hardstem bulrush | SCAC3 | Cyperaceae |
| <i>Schoenoplectus acutus var. acutus</i> | hardstem bulrush | SCACA | Cyperaceae |
| <i>Schoenoplectus americanus</i> | chairmaker's bulrush | SCAM6 | Cyperaceae |

| | | | |
|---|-----------------------|--------------|-------------------------|
| <i>Sisyrinchium idahoense</i> var. <i>occidentale</i> | Idaho blue-eyed grass | SIIDO | Iridaceae |
| <i>Solidago canadensis</i> | Canada goldenrod | SOCA6 | Asteraceae |
| <i>Sophora nuttalliana</i> | silky sophora | SONU | Fabaceae |
| <i>Sporobolus airoides</i> | alkali sacaton | SPAI | Poaceae |
| <i>Sporobolus contractus</i> | spike dropseed | SPCO4 | Poaceae |
| <i>Symphyotrichum falcatum</i> | white prairie aster | SYFA | Asteraceae |
| <i>Symphyotrichum laeve</i> | smooth blue aster | SYLA3 | Asteraceae |
| <i>Symphyotrichum lanceolatum</i> | white panicle aster | SYLA6 | Asteraceae |
| <i>Symphyotrichum lanceolatum</i> | white panicle aster | SYLA6 | Asteraceae |
| <i>Taraxacum officinale</i> | common dandelion | TAOF | Asteraceae |
| <i>Toxicodendron rydbergii</i> | western poison ivy | TORY | Anacardiaceae |
| <i>Tragopogon dubius</i> | yellow salsify | TRDU | Asteraceae |
| <i>Trifolium pratense</i> | red clover | TRPR2 | Fabaceae |
| <i>Trifolium repens</i> | white clover | TRRE3 | Fabaceae |
| <i>Typha latifolia</i> | broadleaf cattail | TYLA | Typhaceae |
| <i>Ulmus pumila</i> | Siberian elm | ULPU | Ulmaceae |
| <i>Verbascum thapsus</i> | common mullein | VETH | Scrophulariaceae |
| <i>Veronica americana</i> | American speedwell | VEAM2 | Scrophulariaceae |
| <i>Vicia americana</i> | American vetch | VIAM | Fabaceae |
| <i>Xanthium strumarium</i> | rough cocklebur | XAST | Asteraceae |

Appendix 3. Field forms for rapid assessment techniques evaluated during pilot work

Field forms for Stream Reach Inventory and Channel Stability Evaluation method (Pfankuch 1975).

| UPPER BANKS | EXCELLENT | GOOD | FAIR | POOR |
|---|---|---|---|--|
| Landform slope | Bank slope gradient <30% | 2 Bank slope gradient 30-40% | 4 Bank slope gradient 40-60% | 6 Bank slope gradient >60% |
| Mass-wasting (existing or potential) | No evidence of post or any potential for future mass-wasting into channel. | 3 Infrequent and/or very small. Mostly healed over. Low future potential. | 6 Moderate frequency and size, with some raw spots eroded by water during high flows. | 9 Frequent or large, causing sediment OR imminent danger of same. |
| Debris jam potential (floatable objects) | Essentially absent from immediate channel area. | 2 Present but mostly small twigs and limbs. | 4 Present, volume and size are both increasing. | 6 Moderate to heavy amounts, mainly larger sizes. |
| Vegetative bank protection | >90% plant density. Vigor and variety suggests a deep, dense, soil binding root mass. | 3 70-90% density. Fewer plant species or lower vigor suggests a less dense or deep root mass. | 6 50-70% density. Lower vigor and species form a somewhat shallow and discontinuous root mass. | 9 <50% density plus fewer species and vigor indicate discontinuous and shallow root mass. |
| Channel capacity | Ample for present plus some increases. Peak flows contained. Width to Depth (W/D) ratio <7. | 1 Adequate. Overbank flows rare. W/D ratio 8 to 15. | 2 Barely contains present peaks. Occasional over-bank floods. W/D ratio 15 to 25. | 3 Inadequate. Overbank flows common. W/D ratio >25. |
| LOWER BANKS | | | | |
| Bank rock content | 65% with large, angular boulders 30cm numerous. | 2 40 to 65%, mostly small boulders to cobbles 15-30cm. | 4 20 to 40%, with most in the 7.5-15cm diameter class. | 6 <20% rock fragments of gravel sizes, 2.5-7.5 cm or less. |
| Obstructions (flow deflectors Sediment traps) | Rocks and old logs firmly embedded. Flow pattern without cutting or deposition. Pools and riffles stable. | 2 Some present, causing erosive cross currents and minor pool filling. Obstructions and deflectors newer and less firm. | 4 Moderately frequent, unstable obstructions and deflectors move with high water causing bank cutting and filling of pools. | 6 Frequent obstructions and deflectors cause bank erosion. Sediment traps' full channel migration occurring. |
| Undercutting | Little or none evident. Infrequent raw banks <150cm high. | 4 Some, intermittently at outcrops and constrictions. Raw banks <30cm. | 8 Significant. Cuts 15-30cm high. Root mat overhangs and sloughing evident. | 12 Almost continuous cuts, some >30cm high. Failure of overhangs |
| Deposition | Little or no enlargement of channel or point bars. | 4 Some new increase in bar formation, mostly from coarse gravels. | 8 Moderate deposition of new gravel and coarse sand on old and some new bars. | 12 Extensive deposits of predominantly fine particles. Accelerated |
| STREAM BED | | | | |
| Rock angularity | Sharp edges and corners, plane surfaces roughened. | 1 Rounded corners and edges. Smooth and flat. | 2 Corners and edges well rounded in two dimensions. | 3 Well rounded in all dimensions. |
| Brightness | Surfaces dull, darkened or stained. Not "bright". | 1 Mostly dull, but may have up to 35% bright surfaces. | 2 Mixture, 50-50% dull and bright i.e. 35-65%. | 3 Predominantly bright, 65%, exposed surfaces. |
| Consolidation or particle packing | Assorted sizes tightly packed and/or overlapping. | 2 Moderately packed with some overlapping. | 4 Mostly a loose assortment with no apparent overlap. | 6 No packing evident. Loose, easily moved. |
| Bottom size distribution & stable | No change in sizes evident. Stable materials 80-100% | 4 Distribution shift slight. Stable materials 50-80%. | 8 Moderate change in sizes. Stable materials 20-50% | 12 Marked change. Stable materials 0-20% |
| Scouring and deposition | <5% of the bottom affected by scouring and deposition. | 6 5-30% affected. Scour at constrictions and where steep. Pool deposition. | 12 30-50% affected. Deposits and scour at obstructions, constrictions, and bends. | 18 > 50% of bed in a state of flux or change nearly year-long. |
| Clinging aquatic vegetation (moss and algae) | Abundant, growth largely moss, dark green, perennial. In swift water too. | 1 Common. Algal forms in low velocity and pool areas. Moss and swifter waters. | 2 Present but spotty, mostly in backwater areas. Seasonal blooms | 3 Perennial types scarce 4 or absent. Yellow-green, short term bloom present. |
| COLUMN TOTALS | | | | |

Field data forms used in the habitat assessment field data protocol.

HABITAT ASSESSMENT FIELD DATA SHEET—LOW GRADIENT STREAMS

| | | |
|-------------------------|-----------------|-----------------------------|
| STREAM NAME _____ | | LOCATION _____ |
| STATION # _____ | REACH ID# _____ | STREAM CLASS _____ |
| UTM N _____ | UTM E _____ | RIVER BASIN _____ |
| STORET # _____ | | AGENCY _____ |
| INVESTIGATORS _____ | | |
| FORM COMPLETED BY _____ | | DATE _____ PM TIME _____ |
| REASON FOR SURVEY _____ | | |

| Parameters to be evaluated in sampling reach | Habitat Parameter | Condition Category | | | |
|--|--|--|---|---|--|
| | | Optimal | Suboptimal | Marginal | Poor |
| | 1. Epifaunal Substrate/ Available Cover | Greater than 50% of substrate favorable for epifaunal colonization and fish cover; mix of snags, submerged logs, undercut banks, cobble or other stable habitat and at stage to allow full colonization potential (i.e., logs/snags that are <u>not</u> new fall and not transient). | 30-50% mix of stable habitat; well-suited for full colonization potential; adequate habitat for maintenance of populations; presence of additional substrate in the form of newfall, but not yet prepared for colonization (may rate at high end of scale). | 10-30% mix of stable habitat; habitat availability less than desirable; substrate frequently disturbed or removed. | Less than 10% stable habitat; lack of habitat is obvious; substrate unstable or lacking. |
| | SCORE | 20 19 18 17 16 | 15 14 13 12 11 | 10 9 8 7 6 | 5 4 3 2 1 0 |
| | 2. Pool Substrate Characterization | Mixture of substrate materials, with gravel and firm sand prevalent; root mats and submerged vegetation common. | Mixture of soft sand, mud, or clay; mud may be dominant; some root mats and submerged vegetation present. | All mud or clay or sand bottom; little or no root mat; no submerged vegetation. | Hard-pan clay or bedrock; no root mat or vegetation. |
| | SCORE | 20 19 18 17 16 | 15 14 13 12 11 | 10 9 8 7 6 | 5 4 3 2 1 0 |
| | 3. Pool Variability | Even mix of large-shallow, large-deep, small-shallow, small-deep pools present. | Majority of pools large-deep; very few shallow. | Shallow pools much more prevalent than deep pools. | Majority of pools small-shallow or pools absent. |
| | SCORE | 20 19 18 17 16 | 15 14 13 12 11 | 10 9 8 7 6 | 5 4 3 2 1 0 |
| | 4. Sediment Deposition | Little or no enlargement of islands or point bars and less than <20% of the bottom affected by sediment deposition. | Some new increase in bar formation, mostly from gravel, sand or fine sediment; 20-50% of the bottom affected; slight deposition in pools. | Moderate deposition of new gravel, sand or fine sediment on old and new bars; 50-80% of the bottom affected; sediment deposits at obstructions, constrictions, and bends; moderate deposition of pools prevalent. | Heavy deposits of fine material, increased bar development; more than 80% of the bottom changing frequently; pools almost absent due to substantial sediment deposition. |
| | SCORE | 20 19 18 17 16 | 15 14 13 12 11 | 10 9 8 7 6 | 5 4 3 2 1 0 |
| | 5. Channel Flow Status | Water reaches base of both lower banks, and minimal amount of channel substrate is exposed. | Water fills >75% of the available channel; or <25% of channel substrate is exposed. | Water fills 25-75% of the available channel, and/or riffle substrates are mostly exposed. | Very little water in channel and mostly present as standing pools. |
| | SCORE | 20 19 18 17 16 | 15 14 13 12 11 | 10 9 8 7 6 | 5 4 3 2 1 0 |

HABITAT ASSESSMENT FIELD DATA SHEET—LOW GRADIENT STREAMS

| Habitat Parameter | Condition Category | | | | | | | | | | | | | | | | | | | | |
|--|--|----|----|----|----|--|----|----|----|----|---|---|---|---|---|---|---|---|---|---|---|
| | Optimal | | | | | Suboptimal | | | | | Marginal | | | | | Poor | | | | | |
| 6. Channel Alteration | Channelization or dredging absent or minimal; stream with normal pattern. | | | | | Some channelization present, usually in areas of bridge abutments; evidence of past channelization, i.e., dredging, (greater than past 20 yr) may be present, but recent channelization is not present. | | | | | Channelization may be extensive; embankments or shoring structures present on both banks; and 40 to 80% of stream reach channelized and disrupted. | | | | | Banks shored with gabion or cement; over 80% of the stream reach channelized and disrupted. Instream habitat greatly altered or removed entirely. | | | | | |
| SCORE | 20 | 19 | 18 | 17 | 16 | 15 | 14 | 13 | 12 | 11 | 10 | 9 | 8 | 7 | 6 | 5 | 4 | 3 | 2 | 1 | 0 |
| 7. Channel Sinuosity | The bends in the stream increase the stream length 3 to 4 times longer than if it was in a straight line. (Note - channel braiding is considered normal in coastal plains and other low-lying areas. This parameter is not easily rated in these areas.) | | | | | The bends in the stream increase the stream length 1 to 2 times longer than if it was in a straight line. | | | | | The bends in the stream increase the stream length 1 to 2 times longer than if it was in a straight line. | | | | | Channel straight; waterway has been channelized for a long distance. | | | | | |
| SCORE | 20 | 19 | 18 | 17 | 16 | 15 | 14 | 13 | 12 | 11 | 10 | 9 | 8 | 7 | 6 | 5 | 4 | 3 | 2 | 1 | 0 |
| 8. Bank Stability (score each bank) | Banks stable; evidence of erosion or bank failure absent or minimal; little potential for future problems. <5% of bank affected. | | | | | Moderately stable; infrequent, small areas of erosion mostly healed over. 5-30% of bank in reach has areas of erosion. | | | | | Moderately unstable; 30-60% of bank in reach has areas of erosion; high erosion potential during floods. | | | | | Unstable; many eroded areas; "raw" areas frequent along straight sections and bends; obvious bank sloughing; 60-100% of bank has erosional scars. | | | | | |
| SCORE (LB) | Left Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |
| SCORE (RB) | Right Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |
| 9. Vegetative Protection (score each bank) Note: determine left or right side by facing downstream. | More than 90% of the streambank surfaces and immediate riparian zone covered by native vegetation, including trees, understory shrubs, or nonwoody macrophytes; vegetative disruption through grazing or mowing minimal or not evident; almost all plants allowed to grow naturally. | | | | | 70-90% of the streambank surfaces covered by native vegetation, but one class of plants is not well-represented; disruption evident but not affecting full plant growth potential to any great extent; more than one-half of the potential plant stubble height remaining. | | | | | 50-70% of the streambank surfaces covered by vegetation; disruption obvious; patches of bare soil or closely cropped vegetation common; less than one-half of the potential plant stubble height remaining. | | | | | Less than 50% of the streambank surfaces covered by vegetation; disruption of streambank vegetation is very high; vegetation has been removed to 5 centimeters or less in average stubble height. | | | | | |
| SCORE (LB) | Left Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |
| SCORE (RB) | Right Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |
| 10. Riparian Vegetative Zone Width (score each bank riparian zone) | Width of riparian zone >18 meters; human activities (i.e., parking lots, roadbeds, clear-cuts, lawns, or crops) have not impacted zone. | | | | | Width of riparian zone 12-18 meters; human activities have impacted zone only minimally. | | | | | Width of riparian zone 6-12 meters; human activities have impacted zone a great deal. | | | | | Width of riparian zone <6 meters; little or no riparian vegetation due to human activities. | | | | | |
| SCORE (LB) | Left Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |
| SCORE (RB) | Right Bank | | 10 | 9 | | 8 | | 7 | 6 | | 5 | | 4 | 3 | | 2 | | 1 | 0 | | |

Parameters to be evaluated broader than sampling reach

PFC checklist field form

Standard Checklist

Name of Riparian-Wetland Area: _____

Date: _____ Segment/Reach ID: _____

Miles: _____ Acres: _____

ID Team Observers: _____

| Yes | No | N/A | HYDROLOGY |
|-----|----|-----|---|
| | | | 1) Floodplain above bankfull is inundated in "relatively frequent" events |
| | | | 2) Where beaver dams are present they are active and stable |
| | | | 3) Sinuosity, width/depth ratio, and gradient are in balance with the landscape setting (i.e., landform, geology, and bioclimatic region) |
| | | | 4) Riparian-wetland area is widening or has achieved potential extent |
| | | | 5) Upland watershed is not contributing to riparian-wetland degradation |

| Yes | No | N/A | VEGETATION |
|-----|----|-----|---|
| | | | 6) There is diverse age-class distribution of riparian-wetland vegetation (recruitment for maintenance/recovery) |
| | | | 7) There is diverse composition of riparian-wetland vegetation (for maintenance/recovery) |
| | | | 8) Species present indicate maintenance of riparian-wetland soil moisture characteristics |
| | | | 9) Streambank vegetation is comprised of those plants or plant communities that have root masses capable of withstanding high streamflow events |
| | | | 10) Riparian-wetland plants exhibit high vigor |
| | | | 11) Adequate riparian-wetland vegetative cover is present to protect banks and dissipate energy during high flows |
| | | | 12) Plant communities are an adequate source of coarse and/or large woody material (for maintenance/recovery) |

| Yes | No | N/A | EROSION/DEPOSITION |
|-----|----|-----|--|
| | | | 13) Floodplain and channel characteristics (i.e., rocks, overflow channels, coarse and/or large woody material) are adequate to dissipate energy |
| | | | 14) Point bars are revegetating with riparian-wetland vegetation |
| | | | 15) Lateral stream movement is associated with natural sinuosity |
| | | | 16) System is vertically stable |
| | | | 17) Stream is in balance with the water and sediment being supplied by the watershed (i.e., no excessive erosion or deposition) |

(Revised 1998)

Stream Visual Assessment (SVA) field form

Stream Visual Assessment Protocol

Owners name _____ Evaluator's name _____ Date _____
Stream name _____ Waterbody ID number _____
Reach location _____

Ecoregion _____ Drainage area _____ Gradient _____
Applicable reference site _____
Land use within drainage (%): row crop _____ hayland _____ grazing/pasture _____ forest _____ residential _____
confined animal feeding operations _____ Cons. Reserve _____ industrial _____ Other: _____
Weather conditions-today _____ Past 2-5 days _____
Active channel width _____ Dominant substrate: boulder _____ gravel _____ sand _____ silt _____ mud _____

Site Diagram

Assessment Scores

| | |
|---------------------------|--------------------------|
| Channel condition | <input type="checkbox"/> |
| Hydrologic alteration | <input type="checkbox"/> |
| Riparian zone | <input type="checkbox"/> |
| Bank stability | <input type="checkbox"/> |
| Water appearance | <input type="checkbox"/> |
| Nutrient enrichment | <input type="checkbox"/> |
| Barriers to fish movement | <input type="checkbox"/> |
| Instream fish cover | <input type="checkbox"/> |

| | |
|----------------------|--------------------------|
| Pools | <input type="checkbox"/> |
| Invertebrate habitat | <input type="checkbox"/> |

| Score only if applicable | |
|--|--------------------------|
| Canopy cover | <input type="checkbox"/> |
| Manure presence | <input type="checkbox"/> |
| Salinity | <input type="checkbox"/> |
| Riffle embeddedness | <input type="checkbox"/> |
| Macroinvertebrates Observed (optional) | <input type="checkbox"/> |

| | | |
|---|---------|-----------|
| Overall score (Total divided by number scored) | <6.0 | Poor |
| | 6.1-7.4 | Fair |
| | 7.5-8.9 | Good |
| | >9.0 | Excellent |

Suspected causes of observed problems _____

Recommendations _____
